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Draft Proposal to Establish a 24-hour Standard for PM_{2.5}

Review of the California Ambient Air Quality Standards For Particulate Matter and Sulfates

Report to the Air Quality Advisory Committee

Public Review Draft

March 12, 2002

California Environmental Protection Agency

**Air Resources Board
and
Office of Environmental Health Hazard Assessment**

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California Environmental Protection Agency

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Draft Proposal to Establish a 24-hour standard for PM2.5

Background

In the initial Report to the Air Quality Advisory Committee (November 30, 2001), Air Resources Board (ARB) and Office of Environmental Health Hazard Assessment (OEHHA) staff did not propose a specific 24-hour standard for PM2.5. The Committee, however, unanimously recommended that staff develop such a standard, and suggested several possible approaches. Responding to the Committee's concerns and suggestions, OEHHA staff members have formulated the following recommendation, in consultation with staff at the ARB.

As reviewed in prior sections, the epidemiological literature suggests the existence of impacts on both morbidity and mortality related to fluctuations in ambient PM2.5 on a daily basis. Morbidity outcomes associated with changes in 24-hour concentrations in PM2.5 include admissions to hospitals for respiratory and cardiac diseases (see sections 7.5.1 and 7.5.2). There is also a growing literature suggesting potential mechanistic linkages between ambient PM2.5 and exacerbations of cardiovascular disease that could result in hospitalization or death (see section 7.8). These include associations with serious cardiac arrhythmias, myocardial infarctions, and decreased heart rate variability (Peters et al., 2000; 2001, Liao et al., 1999; Gold et al., 2000; Pope et al. 1999). As noted in prior sections, the entire spectrum of adverse health outcomes associated with ambient PM2.5, including exacerbations of asthma, emergency room visits, hospitalizations, as well as mortality, occurs within the same general concentration range and also seems to be best described by a linear, non-threshold model. Such a model implies that the level(s) at which adverse effects begin to occur cannot be identified and that there are no abrupt changes in the slope of the dose-response relationship to delineate a "bright line" or threshold.

Consistent observations of health effects associated with low ambient concentrations of fine particles, however, indicate that a short-term PM2.5 standard is required to protect public health. Moreover, while state-wide attainment of the proposed annual PM2.5 standard will result in a reduction of PM2.5 peak concentrations, some areas will be able to attain the annual standard and still experience periods during which 24-hour PM2.5 concentrations associated with increased morbidity and mortality can occur (e.g., during winter inversions accompanied by widespread residential wood combustion). This phenomenon also evidences the need for a short-term standard.

Development of a short-term standard for PM2.5, however, encompasses difficulties similar to those encountered with respect to the 24-hour standard for PM10, largely because the exposure-response relationships examined appear to be linear without clear evidence of a threshold. The linear, nonthreshold model carries implications for the determination of an "adequate margin of safety" specified in the language of the Children's Environmental Health Protection Act.

1 In order to address the lack of a “bright line” in the exposure-response curve,
2 OEHHA staff members propose to reduce the entire distribution of fine particles
3 below reported the levels of distributions consistently associated with adverse
4 health effects. The underlying principle is to reduce not only the mean
5 concentration (represented by the annual average), but specifically the upper tail
6 of the distribution, described by the 98th percentile of the distributions of
7 published studies. In so doing, OEHHA has relied primarily on studies relating
8 fine particle concentrations with daily mortality, the most serious irreversible
9 health impact. As noted above and in section 7.5, associations of PM2.5 with
10 morbidity have been observed to occur within the same concentration range as
11 those linked with increased daily mortality. We have therefore assumed that a
12 standard intended to protect against the occurrence of mortality will also protect
13 against these other important health outcomes.

14 **Methodological Approaches**

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16 In developing this recommendation, OEHHA staff followed several
17 approaches. Specifically, we have: (1) used statistical methods to examine the
18 shape of the exposure-response relationships using two California datasets, and
19 compared the results with those reported for other non-California datasets; (2)
20 tabulated the results of all time-series studies published in English, for which
21 direct PM2.5 monitoring data were available, that have explored associations
22 between low levels of ambient PM2.5 and daily mortality; and (3) examined, with
23 technical assistance from ARB staff, the upper tail of the PM2.5 distribution in
24 California consistent with an annual average of 12 $\mu\text{g}/\text{m}^3$, based on data
25 collected throughout California in 1999 and 2000. Based on the results of these
26 analyses, OEHHA recommends that the 24-hour PM2.5 standard be established
27 at a level of 25 $\mu\text{g}/\text{m}^3$, not to be exceeded. The adoption of the accompanying
28 recommendation for an annual PM2.5 standard of 12 $\mu\text{g}/\text{m}^3$ is an integral
29 component of this proposal. Attainment of the recommended annual standard
30 will help shift the entire PM2.5 distribution to the left, and will influence peak
31 concentrations, as well. However, in itself, the annual average will not fully
32 address the issue of brief (i.e., one to several days) increases in PM2.5 levels.
33 Thus, the 24-hour standard is intended to protect Californians against significant
34 short-term elevations of PM2.5.

35 **1. Statistical approaches**

36
37 With the objective of further examining the validity of the linear model
38 between mortality and PM2.5, staff from OEHHA and the Bay Area Air Quality
39 Management District (BAAQMD) undertook a variety of detailed analyses of data
40 from the two published California studies involving 24-hour measurements of
41 PM2.5 and daily mortality counts (in Coachella Valley [Ostro et al., 2000] and
42 Santa Clara County [Fairley, 1999]). The modeling techniques used for the
43 exposure-response functions included piecewise linear regression (e.g., utilizing
44 several “hockey-stick” models), locally weighted smoothing in generalized
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46

additive models, trimming analysis (selectively deleting days with high PM_{2.5} values), and Bayesian models (comparing the likelihoods of various thresholds) to explore the evidence for a nonlinear exposure-response at low PM_{2.5} concentrations. In general, within the concentration range of interest for PM_{2.5}, nonlinear models (and, in particular, models intended to identify possible thresholds) offered no improvement over a linear, nonthreshold model in fitting the data. These analyses, which are not presented in this document, are consistent with results reported by almost all other researchers (except, e.g., for Smith et al., 2000) using datasets from locations outside California. At least for mortality, others have also found that a linear nonthreshold model best characterizes the relationship between ambient PM_{2.5} and adverse health outcomes (Pope, 2000 and section 7.3.5). A corollary of this observation is that, in order to calculate a short-term PM_{2.5} standard, additional information (such as the distributions of PM_{2.5} concentrations in published studies examining exposure-response relationships) may be required.

2. Distributions of PM_{2.5} in daily mortality studies.

OEHHA staff obtained data from the authors of all recently published studies examining ambient PM_{2.5} concentrations in relation to daily nonaccidental mortality. Table 7.a provides information on the estimated percentage change in daily mortality associated with a 10 µg/m³ change in PM_{2.5}. All the point estimates of this relationship in Table 7.a are positive, though not all are statistically significant. The upper tail of the PM_{2.5} distribution in each of these investigations is indicated by the 98th percentile, which is somewhat less subject to the factors determining the most extreme values. Examination of the PM_{2.5} levels in Table 7.a indicates that, when the 98th percentiles of the fine particle distributions are ≤32 µg/m³, and the mean fine particle concentrations are <13 µg/m³, the results are characterized by greater uncertainty, since the confidence intervals for the percent change in mortality include zero. These were studies conducted in Portage (WI), Topeka (KS), and in four Canadian cities (Calgary, Edmonton, Ottawa, and Winnipeg). One partial exception to this observation is Vancouver, British Columbia, which had a 98th percentile PM_{2.5} concentration of 30 µg/m³, though the mean concentration was 13 µg/m³. These results do not imply an absence of effects when peak PM_{2.5} concentrations are below 30 µg/m³; rather, these estimates may be subject to greater uncertainty potentially ascribable to several factors, including fewer health impacts associated with exposure to lower concentrations, exposure measurement error, confounding by co-pollutants or meteorological factors, differences in the composition of particle mixtures, decreased statistical power, and reduced variance in the PM_{2.5} values in studies with lower means. The last explanation is unlikely, however, as we examined the coefficients of variation in the studies with relatively low PM_{2.5} mean concentrations and found that they were generally similar to those in the studies with higher mean levels. In contrast, statistical power (i.e., the ability to detect statistically a real relationship between two variables) is likely to be reduced at lower ambient pollutant

1 concentrations. Based on model simulations conducted by staff at the BAAQMD,
2 the increased uncertainty between lower-level PM2.5 concentrations and daily
3 mortality may be attributable in part to insufficient statistical power.

4
5 Published studies provide some guidance for an appropriate reduction in
6 the distribution of PM2.5. An annual PM2.5 standard of 12 $\mu\text{g}/\text{m}^3$ would
7 represent a level lower than the long-term means of all the studies in which
8 significant associations with changes in daily mortality have been identified (see
9 Table 7.a and section 7.3, above). Attainment of the annual average would, as
10 previously noted, result in an across-the-board reduction of PM2.5, including
11 peak concentrations. Setting a 24-hour standard level below 30 $\mu\text{g}/\text{m}^3$ would
12 shift the upper extreme of the PM2.5 distribution to a level lower than those
13 identified in the studies described above. Because the exposure-response
14 relationship is characterized by a linear, nonthreshold model, such a 24-hour
15 standard does not imply total elimination of health risks when this standard is
16 attained. However, reduction of peak PM2.5 concentrations below those
17 observed in studies reporting adverse effects represents a rational approach to
18 reduce the risk of short-term PM2.5-associated mortality and morbidity and to
19 position the entire distribution of PM2.5 below those for which there is current,
20 published evidence of health effects.

21 22 3. Relationship of Recommended Annual PM2.5 Standards and 24-hour PM2.5 23 Concentrations in California

24 As discussed in Chapter 6, the ARB uses the Expected Peak Day
25 Concentration (EPDC) to determine the “design value” for 24-hour standards.
26 The development of the EPDC uses a statistical model of the highest 20% of the
27 daily values from the previous three years, making it relatively robust with respect
28 to fluctuations in daily meteorological conditions. Specifically, the index will not
29 be unduly influenced by any single day, and exceptional events such as forest or
30 urban fires can be excluded. We used a modified version of this process to
31 examine the upper tail of the PM2.5 distribution (98th percentile) rather than the
32 most extreme values within California. With assistance from ARB staff, we
33 conducted an analysis to determine the relationship between the 98th percentile
34 of the PM2.5 distribution in California and the proposed annual average of 12
35 $\mu\text{g}/\text{m}^3$. This analysis identified the 98th percentile concentrations consistent with
36 an annual average of 12 $\mu\text{g}/\text{m}^3$, given recent statewide distributions of PM2.5.

37
38 Using data from 54 sites around the state, located principally in large
39 urban areas, a linear regression model was performed (linear models fit the data
40 better than non-linear models) relating the 98th percentile of the PM2.5
41 distribution to the annual average for the years 1999 and 2000 for each site. The
42 regression model generated an r^2 of 0.79 and indicated that statewide, the 98th
43 percentile for the distribution of PM2.5 associated with a 12 $\mu\text{g}/\text{m}^3$ annual
44 average is approximately 39 $\mu\text{g}/\text{m}^3$. For sites within the jurisdiction of the South
45 Coast Air Quality Management District, representing the most heavily populated
46 air basin in the state, the predicted 98th percentile concentration is approximately

1 37 $\mu\text{g}/\text{m}^3$, while the corresponding value for three other major air basins (the San
2 Francisco Bay Area, San Joaquin Valley, and Sacramento) is 45 $\mu\text{g}/\text{m}^3$, and that
3 for the South Central Coast is 33 $\mu\text{g}/\text{m}^3$.

4
5 This approach to identify ambient PM_{2.5} 98th percentile concentrations
6 consistent with attainment of the proposed annual average indicates that, at least
7 in some of the heavily populated air basins, predicted concentrations of PM_{2.5}
8 could fall within ranges previously reported to be associated with increased daily
9 mortality (Table 7.2) and morbidity. This modified EPDC exercise suggests the
10 need for a lower short-term standard to limit excursions of PM_{2.5} to protect
11 against increased risks of morbidity and mortality.

12 13 **Recommendation for 24-hour PM_{2.5} Standard**

14
15 Examining the evidence described above, OEHHA recommends that the
16 24-hour PM_{2.5} standard be 25 $\mu\text{g}/\text{m}^3$, not to be exceeded. The rationale for this
17 recommendation is as follows:

18
19 (i) Multiple analyses of the exposure-response relationships between
20 PM_{2.5} and mortality indicate that the data can be fitted most parsimoniously with
21 linear, nonthreshold models. Given the apparent linearity of the exposure-
22 response relationships in the epidemiological data, it is difficult to determine at
23 what concentrations within the PM_{2.5} distributions in each study adverse health
24 effects begin. Intuitively, one would expect greater biological responses and
25 larger numbers of adverse events occurring at higher concentrations, everything
26 else being equal. Nonetheless, in a linear exposure-response relationship,
27 effects may be observed at lower levels as well. (Schwartz et al., 1996)

28
29 The importance of the linear, nonthreshold exposure-response
30 relationship cannot be overemphasized in light of legislation requiring that
31 ambient air quality standards be “established at levels that adequately protect the
32 health of the public, including infants and children, with an adequate margin of
33 safety.” (California Health & Safety Code Section 39606(d)(2)) If a threshold in
34 the exposure-response curve cannot be identified, then specification of an
35 “adequate margin of safety” becomes challenging. The approach OEHHA staff
36 members have adopted in pursuit of this objective has therefore been to: (1)
37 identify indicators of the distribution of PM_{2.5} (specifically the means and 98th
38 percentiles) in epidemiological studies that demonstrate the relationship of
39 ambient fine particles with adverse health impacts, (2) recommend that the
40 distribution of PM_{2.5} in California be reduced below the levels of these
41 distributions, and (3) incorporate a margin of safety in the form of a standard “not
42 to be exceeded”, which will assure that the extreme values of the PM_{2.5}
43 distribution in California will be lower (and in general substantially lower) than the
44 98th percentiles of PM_{2.5} distributions in published studies.

1 (ii) Without placing a short-term limitation on PM_{2.5} concentrations, recent
2 experience in California indicates that even attainment of the recommended
3 annual standard of 12 µg/m³ will allow for excursions well into the range in which
4 adverse effects, including mortality, have been identified in epidemiological
5 studies. Notably, the modified EPDC analysis undertaken by the ARB staff
6 indicates that for several large air basins, the estimated 98th percentile of the
7 PM_{2.5} distribution consistent with attainment of an annual standard of 12 µg/m³
8 would be in excess of 40 µg/m³. Thus, adoption of a 24-hour standard of 25
9 µg/m³ would be intended to limit such excursions.

10
11 (iii) As with PM₁₀, morbidity and mortality outcomes appear to occur
12 within the same PM_{2.5} concentration ranges (See Section 7.5). Therefore, we
13 have focused on mortality as the most serious adverse health outcome. Changes
14 in ambient air quality sufficient to protect against increases in mortality should, *a*
15 *fortiori*, protect against the occurrence of morbidity, too.

16
17 (iv) Among studies examining PM_{2.5} and mortality, the long-term mean
18 concentrations of those finding a significant association varied from 13 to 21
19 µg/m³, while the 98th percentiles of the distributions ranged from 30 to 51 µg/m³.
20 Shifting the entire PM_{2.5} distribution downwards and limiting short-term
21 excursions should reduce the likelihood of fine particle-associated mortality and
22 morbidity. Recommending an annual average of 12 µg/m³ addresses the issue
23 of shifting the overall distribution downwards. By the same token, recommending
24 a 24-hour PM_{2.5} limit of 25 µg/m³ would place the upper extreme of the
25 distribution lower than the 98th percentile of those identified in studies finding
26 significant associations with mortality, thereby incorporating a margin of safety.
27 More specifically, except for the study of Vancouver (Burnett et al., 2000), all
28 published investigations of PM_{2.5} and mortality in which statistically significant
29 effects were detected had 98th percentile PM_{2.5} concentrations of 32 µg/m³ or
30 greater. Positioning the upper extreme of the PM_{2.5} distribution in California at
31 25 µg/m³ effectively incorporates a margin of safety into this recommendation,
32 based on the best available scientific evidence.

Table 7.a: Distributions and Associations of 24-hour PM_{2.5} with Daily Mortality in U.S. and Canadian Cities with Long-term Mean PM_{2.5} Concentrations < 25 µg/m³, Sorted by Reported 98 percentile Concentrations*

City	Study Period	Reference	Mean (µg/m ³)	98th percentile	% Increase (95% CI) per 10µg/m ³
Edmonton	1986-1996	Burnett et al., 2000	10	28	2.18(-1.74, 6.10)
Calgary	1986-1996	Burnett et al., 2000	10	29	0.63(-3.58, 4.84)
Winnipeg	1986-1996	Burnett et al., 2000	10	29	0.38(-3.15, 3.91)
Vancouver	1986-1996	Burnett et al., 2000	13	30	2.56(0.23, 4.89)
Topeka, KS	1979-1988	Schwartz et al., 1996	12	31	0.80(-0.20, 3.60)
Phoenix, AZ	1995-1997	Mar et al., 2000	13	32	2.22(0.00, 5.56)
Portage, WI	1979-1987	Schwartz et al., 1996	11	34	1.20(-0.30, 2.80)
Ottawa	1986-1996	Burnett et al., 2000	12	35	2.45(-0.53, 5.43)
Coachella Valley, CA	1995-1998	Ostro et al., 2000	17	38	4.44(0.00, 8.89)
Toronto	1986-1996	Burnett et al., 2000	15	41	0.91(-0.05, 1.87)
Boston, MA	1979-1986	Schwartz et al., 1996	16	42	2.20(1.50, 2.90)
Windsor	1986-1996	Burnett et al., 2000	18	43	5.20(2.24, 8.16)
Montreal	1984-1993	Goldberg et al., 2001	18	43	1.93(1.16, 2.71)
Kingston	1980-1987	Schwartz et al., 1996	21	44	1.40(0.20, 2.60)
St. Louis, MO	1979-1987	Schwartz et al., 1996	19	46	1.10(0.40, 1.70)
Santa Clara, CA	1990-1996	Fairley, 1999	13	51	3.18(0.00, 6.10)
Montreal	1986-1996	Burnett et al., 2000	15	51	1.23(0.11, 2.35)
Detroit, MI	1992-1994	Lippmann et al., 2000	18	55	1.24(-0.26, 2.83)

*Some data in Table 7.a, particularly most of the 98th percentile values, were obtained directly from the authors of the published reports.

References

- Burnett RT, Brook JR, Dann T, Delocla C, Philips O, Calmak S et al . (2000). Association between particulate- and gas-phase components of urban air pollution and daily mortality in eight Canadian cities. In: Grant LD, ed. PM2000: Particulate Matter and Health. Inhal Toxicol 12(Suppl. 4):15-39.
- Fairley D (1999). Daily mortality and air pollution in Santa Clara County, California: 1989-1996. Environ Health Perspect 107(8):637-41.
- Gold DR, Litonjua A, Schwartz J, Lovett E, Larson A, Nearing B et al. (2000). Ambient pollution and heart rate variability. Circulation 101(11):1267-73.
- Goldberg MS, Burnett RT, Bailar JC III, Tamblyn R, Ernst P, Flegel K et al. (2001). Identification of persons with cardiorespiratory conditions who are at risk of dying from the acute effects of ambient air particles. Environ Health Perspect 109(Suppl. 4):487-94.
- Liao D, Creason J, Shy C, Williams R, Watts R, Zweidinger R (1999). Daily variation of particulate air pollution and poor cardiac autonomic control in the elderly. Environ Health Perspect 107(7):521-5.
- Lippmann M, Ito K, Nadas A, Burnett RT (2000). Association of particulate matter components with daily mortality and morbidity in urban populations. Res Rep Health Eff Inst (95):5-72, discussion 73-82.
- Mar TF, Norris GA, Koenig JQ, Larson TV (2000). Associations between air pollution and mortality in Phoenix, 1995-1997. Environ Health Perspect 108(4):347-53.
- Ostro BD, Broadwin R, Lipsett MJ (2000). Coarse and fine particles and daily mortality in the Coachella Valley, California: a follow-up study. J Expo Anal Environ Epidemiol 10(5):412-9.
- Peters A, Dockery DW, Muller JE, Mittleman MA (2001). Increased particulate air pollution and the triggering of myocardial infarction. Circulation 103(23):2810-5.
- Peters A, Liu E, Verrier RL, Schwartz J, Gold DR, Mittleman M et al. (2000). Air pollution and incidence of cardiac arrhythmia. Epidemiology 11(1):11-7.
- CA III (2000). Epidemiology of fine particulate air pollution and human health: biologic mechanisms and who's at risk? Environ Health Perspect 108 Suppl 4:713-23.
- Pope CA III, Verrier RL, Lovett EG, Larson AC, Raizenne ME, Kanner RE et al. (1999). Heart rate variability associated with particulate air pollution. Am Heart J 138(5 Pt 1):890-9.

- 1 Schwartz J, Dockery DW, Neas LM (1996). Is daily mortality associated specifically
- 2 with fine particles? J Air Waste Manag Assoc 46:927-39.
- 3